Survey on heavy metals contaminated soils in Thai Nguyen and Hung Yen provinces in Northern Vietnam

Khảo sát đất ô nhiễm kim loại nặng ở tỉnh Thái Nguyên và tỉnh Hưng Yên thuộc miền Bắc Việt Nam

Research Article

Chu, Thi Thu Ha*

Institute of Ecology and Biological Resources, Vietnam Academy of Science and Technology, 18 Hoang Quoc Viet, Hanoi, Vietnam

In Vietnam, soil contamination with lead and cadmium at very high level was investigated and discovered in the surrounding areas of zinc-lead mining and processing factory in Tan Long (Dong Hy district, Thai Nguyen province) and around the lead-recycling smelter in Chi Dao (Van Lam district, Hung Yen province). The survey on soil contaminated by arsenic due to the tin mining and sifting activities in Ha Thuong (Dai Tu district, Thai Nguyen province) was also carried out. In Tan Long, the concentrations of lead and cadmium in the old solid waste dump from zinc-lead factory varied from 1,100 to 13,000 mg kg⁻¹, and from 11.34 to 61.04 mg kg⁻¹, respectively. Soil Pollution Indexes (SPI) of lead and cadmium were highest in the old solid waste dump area, followed by the ones in the rice paddy soils. In Chi Dao, the soils of many sites were polluted with lead and cadmium such as in the gardens of lead-recycling households where the concentrations of lead and cadmium were 7,000 - 15,000 mg kg⁻¹ and 1.8 - 3.6 mg kg⁻¹. In rice paddies, the soils were also polluted by lead. SPI of lead in paddy soil areas within 300 m radius from the lead smelter were from 3.6 to 100 fold higher than the safe limit. The sediment from the ditch near the lead smelters contained extremely high levels of lead (7,000 - 110,000 mg kg⁻¹) and cadmium (3.8 - 17.7 mg kg⁻¹). The tin mining and sifting activities in Ha Thuong was the cause for the arsenic contamination of the soil in this area. The arsenic contents in soils at all locations investigated were higher than 320 mg kg⁻¹ (dry weight) and up to 3,809 mg kg⁻¹.

Keywords: heavy metal, metal mining, lead recycling, soil pollution, Thai Nguyen, Hung Yen

* Corresponding author
E-mail: hachuthi@yahoo.com

http://www.openaccess.tu-dresden.de/ojs/index.php/jve/ ISSN: 2193-6471
1. Introduction

Heavy metals contamination of soils has become one of the most serious environmental problems today. The sources of heavy metals in those soils are diverse, including burning of fossil fuels, mining and smelting of metalliferous ores, municipal wastes, fertilisers, pesticides, sewage sludge amendments and the use of pigments and batteries. Some of the heavy metals are micronutrients necessary for plant growth, such as Zn, Cu, Mn, Ni and Co, while others have no known function, such as Cd, Pb and As.

Due to their immutable nature, these heavy metals are a group of pollutants of much concern. Once heavy metals penetrate into the food chain, they can damage the health and life of animals and human because of their toxicities. Lead (Pb) is a kind of heavy metal with neuro-virulent properties that animals and human are very sensitive to. The toxic effect of Pb is mainly based on its ability to react with functional groups such as sulphydryl, carboxyl and amine, leading to a decrease or loss of activity of many enzymes that are important for cell functions (Peng et al., 2005). People living around Pb-Zn mining areas in Thai Nguyen province of Vietnam occurred some sign of heavy metal poisoning even cancer disease (Gia Lai Online, 2011). And in a lead recycling village in Hung Yen province of Vietnam, a lot of children became less intelligent than the children in other places of this province (Viet Bao, 2011).

Cadmium (Cd) is an occupational and environmental pollutant implied in the development of various types of cancer. It is carcinogenic in animals under certain exposure conditions and may enhance the occurrence of lung and possibly of prostate cancer in workers exposed to high airborne concentrations (Lawerys et al., 1990). In fact, there were a lot of people poisoned by Cd (several deaths were also reported) in cases such as the Cd pollution in Toyama prefecture, Japan, in 1955 (Fleischer et al., 1974). The disease was epidemic among elderly women who gave birth to many children. The outstanding features of the disease were lumbar pain and myalgia, spontaneous fractures with skeleton deformation.

Arsenic (As) is a highly bioactive and toxic element, its presence at elevated levels in soils and drinking water being threatening food safety and human health. It adversely affects biological activities as a teratogen, carcinogen or mutagen as well as having detrimental effects on the digestive system, respiratory system and immune system (Sun et al., 2007). In Vietnam, arsenic caused some people have pain and injury in heart, lung, even death in the case of a 7 year-old child in 2001 (VnExpress, 2011). Inorganic arsenic (arsenate and arsenite) is highly toxic to plants because it uncouples phosphorylation and inhibits phosphate uptake. At higher concentrations, arsenic interferes with plant metabolic processes and can inhibit growth and under severe conditions may lead to plant’s death (Geng et al., 2006).

2. Location, materials and methods

2.1 Study area

Thai Nguyen and Hung Yen are two provinces located in the north of Vietnam and near Hanoi (Figure 1).

![Figure 1. Location of Thai Nguyen and Hung Yen provinces](image1)

The Zn/Pb mining operation areas in Tan Long (Dong Hy district, Thai Nguyen province), Pb smelter areas in Chi Dao (Van Lam district, Hung Yen province) and Sn-mining operation areas in Ha Thuong (Dai Tu district, Thai Nguyen province) of Vietnam are the places focused on for investigating the Pb, Cd, As contents in soils.

Tan Long is the place where a former Zn/Pb mining and processing French factory operated in the early of 20th century. After 1954, the activities were ceased but since 1980, they were re-started. The disposal of the waste from mining and processing of heavy metals caused Cd and Pb pollution in the surrounding environment (Figure 2).

![Figure 2. The new waste dump disposed by the Zn/Pb factory in Tan Long](image2)

Chi Dao is the village whose population dealt with the Pb-recycling since 1978 up to now. The people collected many old batteries to take the Pb for the recycling process (Figure 3). Pb and Cd from the smelters was eliminated into atmosphere, soil and water and might have effects on the vegetable and rice products of the area.
2.2 Materials and methods

Soil samples were collected at the surface zone within 0 - 20 cm. They were air-dried at room temperature for 5 - 7 days. The large-size debris, stones and pebbles were removed, then the soil was grinded and sieved through a 2 mm polyethylene sieve. The mineralization of samples was done with HNO₃ 65% (Merck-Germany) and some droplets of H₂O₂ at 100 - 110 °C. The Pb and Cd contents were analyzed by dithizone and Atom Absorption Spectroscopy (AAS) methods. The As content was analyzed by AAS method.

The accuracy of the analysis process is controlled by blank, duplication and reference samples (Tort-2 of Canada). The heavy metal pollution levels in soil are assessed by the Soil Pollution Indexes (SPI):

$$SPI = \frac{\text{Heavy metal content in study soil}}{\text{Safe level (Vietnam standard)}} \times 100$$

SPI<100: bellow the pollution level; SPI=100: safe level; SPI>100: pollution level.

3. Results and discussion

3.1 Soil pollution in Tan Long

At the site near the Zn/Pb mining and processing factory in Tan Long, solid wastes of ore-sifting operation are going through gutter to the dumps near the factory. The source of heavy metal pollution in soil is the wastewater and the solid wastes of ore-sifting activities, as well as the dust from ore grinding, transportation and storing. There are two waste dump areas: the old one on the way to the factory (abundant long time ago) and the new one which is about 1 km far from the factory which several years ago received the waste flow of the factory daily (Figure 2).

For assessing the pollution level of interested heavy metals in this area, hundreds of soil samples were collected and analyzed. The analysis results of the soil samples collected at six main sites are shown in Table 1. For the site containing old waste dump, Pb and Cd contents were the highest (1,100 - 13,000 mg·kg⁻¹ and 11.34 - 61.04 mg·kg⁻¹, respectively). This was followed by the new waste dump area with concentrations of Pb and Cd ranging from 5,300 - 9,200 mg·kg⁻¹ and 5.90 - 9.05 mg·kg⁻¹, respectively. In the paddy soils, the concentrations of Pb and Cd were 1,271 - 3,953 mg·kg⁻¹ and 2.30 - 42.90 mg·kg⁻¹. The garden soils near the new waste dump had contents of Pb and Cd at the lowest levels (27.9 - 35.8 mg·kg⁻¹ and 0.08 - 0.12 mg·kg⁻¹, respectively). The garden soils near the old waste dump had the Pb content of all study sites and the Cd content of some collected samples higher than the safe levels recommended by the Vietnam standard 7209:2002 (Pb: 230 - 360 mg·kg⁻¹ in comparison with 70 mg·kg⁻¹ and Cd: 0.6 - 3.4 mg·kg⁻¹ in comparison with 2.0 mg·kg⁻¹).

Concerning the Soil Pollution Indexes (SPI), the SPI values of Pb and Cd in the soil of old waste dump were: 1,570 - 18,570 and 570 - 3,050, respectively. These results showed that the soils in the old waste dump had the Pb content exceeding the safe level from 15.7 to 185.7 times, and the Cd content exceeding the safe level from 5.7 to 30.5 times (Vietnam standard).

In the paddy soils near the old waste dump, the SPI of Pb and Cd were 1,810 - 5,650 and 115 - 2,150. In the garden soils near the old waste dump, the SPI of Pb and Cd were 390 - 510 and 30 - 170, higher than the safe levels for Pb by 3.9 - 5.1 fold, and in some samples, the concentrations of Cd exceeded the safe level by up to 1.7 fold. The amount of heavy metals in the soils at this area mainly originated from the residues accumulated long time ago and from precipitations since this waste dump no longer received the waste from the Zn/Pb factory. The soils near the new waste dump had the SPI-Pb of 230 to 1,290, exceeding the safe level by 2.3 - 12.9 times. However, the soils were not polluted by Cd (SPI-Cd: 6 - 70).

The garden soils near the new waste dump were just slightly contaminated by these two heavy metals with very low SPI values (SPI-Pb: 40 - 50, SPI-Cd: 4 - 6). These low indexes proved that the background soils in this area were very poor in Pb and Cd. The waste of Zn/Pb factory contained a large amount of Pb and Cd, this being the source of toxic heavy metal pollution that needed to be controlled. From the new waste dump of the Zn/Pb factory, the heavy metals were eliminated every day and this led to the Pb and Cd contamination of the adjacent soil areas at higher level than in garden soils. Thus, in Tan Long, around the Zn/Pb factory, soils were polluted by Pb and Cd due to the mining activities, the old waste dump area had a higher pollution level than the new waste dump.
3.2 Soil pollution in Chi Dao

In Chi Dao, Pb-recycling activities have been operated since 1978 at many households in the village and at the centralized smelter area in the paddy field near the road. The Pb was released into soil from three main sources: wastewater of the battery breaking, solid waste and dust smoke of recycling processes.

Table 2 shows that, in the village, the garden soils of Pb-recycling households contained Pb and Cd concentrations (7,000 - 15,000 mg·kg⁻¹ and 1.8 - 3.6 mg·kg⁻¹, respectively) much higher than the garden soils of the household that were not affected by Pb-recycling activities (40 - 60 and 0.5 - 0.8 mg·kg⁻¹, respectively). In the paddy field, the Pb and Cd contents in the soil less than 20 m away from smelters were 2,000 - 10,000 mg·kg⁻¹ and 2.5 - 56.5 mg·kg⁻¹, respectively; in the paddy soil about 50 m away from the smelters the concentrations ranged from 1,100 to 7,000 mg·kg⁻¹ and from 0.5 to 1.3 mg·kg⁻¹. In the paddy soil 100 m away from smelters the concentrations were 950 - 3,600 mg·kg⁻¹ (Pb) and 0.2 - 1.5 mg·kg⁻¹ (Cd). In the paddy soil more far away from smelters (200 - 300 m), the concentrations were 250 - 770 mg·kg⁻¹ for Pb and 0.2 - 1.2 mg·kg⁻¹ for Cd. In general, the Pb contents in soil decreased with the distance from the smelters but the Cd content did not always follow the same trend.

The cultivated paddy soil area near the road was used for control. In this case, the assessment method was similar to that used for the soil areas near the smelters. The Pb and Cd contents were 55 - 90 mg·kg⁻¹ and 0.2 - 1.0 mg·kg⁻¹, these values being higher than those in the garden soils of the households that were not affected by Pb recycling. This situation can be explained by the effects of traffic activities and agricultural productions. In comparison with the control areas, all the soil samples collected at the paddy soil areas located 20 - 300 m far from the smelter had higher concentrations of both heavy metals investigated. This proved that Pb-recycling process caused the increasing of Pb and Cd amount in the soil of the surrounding areas.

Table 2. Contents in soils and SPI of Pb and Cd at study sites in Chi Dao

<table>
<thead>
<tr>
<th>No</th>
<th>Sites</th>
<th>Contents (dry weight, mg·kg⁻¹)</th>
<th>Soil Pollution Indexes (SPI)</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td></td>
<td>Pb</td>
<td>Cd</td>
</tr>
<tr>
<td>1</td>
<td>Soil in village</td>
<td></td>
<td></td>
</tr>
<tr>
<td></td>
<td>- Garden soils of Pb-recycling household</td>
<td>7,000 - 15,000</td>
<td>1.8 - 3.6</td>
</tr>
<tr>
<td></td>
<td>- Garden soils of household which are not affected by Pb recycling</td>
<td>40 - 60</td>
<td>0.5 - 0.8</td>
</tr>
<tr>
<td>2</td>
<td>Soil area less than 20 m away from smelters</td>
<td>2,000 - 10,000</td>
<td>2.5 - 56.5</td>
</tr>
<tr>
<td>3</td>
<td>Paddy soil 50 m away from smelters</td>
<td>1,100 - 7,000</td>
<td>0.5 - 1.3</td>
</tr>
<tr>
<td>4</td>
<td>Paddy soil 100 m away from smelters</td>
<td>950 - 3,600</td>
<td>0.2 - 1.5</td>
</tr>
<tr>
<td>5</td>
<td>Paddy soil 200-300 m away from smelters</td>
<td>250 - 770</td>
<td>0.2 - 1.2</td>
</tr>
<tr>
<td>6</td>
<td>Control paddy soil area</td>
<td>55 - 90</td>
<td>0.2 - 1.0</td>
</tr>
<tr>
<td>7</td>
<td>Vietnam standard 7209:2002</td>
<td>70</td>
<td>2.0</td>
</tr>
<tr>
<td>8</td>
<td>Sediment of ditch</td>
<td>7000 - 110,000</td>
<td>3.8 - 17.7</td>
</tr>
</tbody>
</table>

In the water ditch near the smelter area (which was continuously receiving the waste from the Pb-recycling process), the Pb and Cd contents in sediment were very high (Pb: 7,000 – 110,000 mg·kg⁻¹, Cd: 3.8 – 17.7 mg·kg⁻¹). This was a dangerous situation because heavy metals could be transferred and accumulated in the paddy soil through dredging and amendment to the field, where the rice and other crops were cultivated. The SPI values showed that the soil in areas affected by Pb-recycling activities was heavily polluted by Pb. The paddy soil around 300 m away from smelter areas had the SPI from 360 to 10,000 that is 3.6 - 100 times higher than the safe level. In case of Cd, only the soil area less than 20 m around smelters was polluted.

3.3 Soil pollution in Ha Thuong

In Ha Thuong, due to Sn-mining operation, there were many fallow soil areas. These were old mine areas (hamlet 4) and paddy soil areas affected by waste
4. Conclusions

flowing from ore sift site nearby (hamlet 7), or from mining and sifting area (hamlet 6). The most typical characteristics of the soils in these areas were the low pH values (3.5 – 4.5, data not shown), the high content in As and Pb but with Cd content lower than 1 mg kg\(^{-1}\) (data not shown). Data from Table 3 show that, among the study sites, the highest As content was found in the grass soil of hamlet 7 (841 – 3,809 mg kg\(^{-1}\)), followed by the grass soil of hamlet 4 (327 – 1,683 mg kg\(^{-1}\)), while the lowest value was measured in the grass soil of hamlet 6 (350 – 592 mg kg\(^{-1}\), dry weight).

Table 3. Contents in soils and SPI of As at study sites in Ha Thuong (Dang and Chu, 2009)

<table>
<thead>
<tr>
<th>N°</th>
<th>Sites</th>
<th>Contents (dry weight, mg kg(^{-1}))</th>
<th>Soil Pollution Indexes (SPI)</th>
</tr>
</thead>
<tbody>
<tr>
<td>1</td>
<td>Grass soil area in hamlet 4</td>
<td>327 – 1,683</td>
<td>2720 – 14,020</td>
</tr>
<tr>
<td>2</td>
<td>Grass soil area in hamlet 6</td>
<td>350 – 592</td>
<td>2,920 – 4,930</td>
</tr>
<tr>
<td>3</td>
<td>Grass soil area in hamlet 7</td>
<td>841 – 3,809</td>
<td>7,010 – 31,740</td>
</tr>
<tr>
<td>4</td>
<td>Vietnam standard 7209:2002</td>
<td>12</td>
<td>100</td>
</tr>
</tbody>
</table>

In fact, the grasses and wild plants that grow at these sites might be a food source for the animals. Therefore, even if these areas are not used for crops cultivating, the potential adverse effects on human and animals are still of concern because the SPI proved that all the study sites were seriously contaminated with As.

In the world, many research results showed similar cases of heavy metals pollution. For example, in China there are over 8,000 national and 230,000 private mining companies operating, resulting in 200,000 km\(^2\) of derelict land, which includes the loss of 370,000 km\(^2\) of agricultural land (Young, 1998). In a national survey commissioned by the United Kingdom, the Department of the Environment analysed samples collected from November 1981 to June 1982 from 53 locations in England, Scotland, and Wales (Thornton et al., 1990). The results showed that 93 % of the garden soils exceeded 2,000 mg kg\(^{-1}\) Pb. In Derbyshire, the Pb concentration was within the range of 1,180 – 22,100 mg kg\(^{-1}\), while the Pb concentration in the vegetable plot soil the range was 1,140 – 26,500 mg kg\(^{-1}\).

Several studies reported that the Cd content in soils near the smelter or metallurgical factories were very high. For example, in Poland, at a site located 600 m away from a Zn metallurgical factory, the Cd content in soil was 250 mg kg\(^{-1}\) (Greszta and Godzik, 1969). Belgium is an important producer of Cd (about 25% of the European production). Certain areas of the country are also polluted by Cd, mainly due to past emissions from non-ferrous industries. Cd concentration in airborne particles, soil and grass increased through years from 1970s to 1980s (Knetzschmar et al., 1980).

In Vietnam, several scientists investigated the heavy metal contents in soil in several areas (Pham, 2002; Pham et al., 2002; Nguyen, 2003; Tran and Tran, 2002). They found that heavy metals levels were under the safe level according to the Vietnamese standard. However, by using wastewater for the irrigation of crops, the land might have been exposed to heavy metals contamination (Nguyen et al., 2001). According to Le and Nguyen (1998) and to Ho and Nguyen (2003), most of the soil pollutions with Pb and Cu are caused by the traditional craft villages and heavy metal recycling activities.

4. Conclusions

In Tan Long, the concentrations of Pb and Cd in soil samples at almost of all the investigated sites exceeded the limitation level according to the Vietnamese standard, except for the Pb and Cd concentrations in the garden soils near the new waste dump and Cd concentration in the soil area adjacent to the new waste dump. In Chi Dao, the Pb contents in almost of all the soil samples at the investigated sites exceeded the safe level. For Cd, the soils collected from two sites (garden soils of Pb-recycling household and soil area less than 20 m far from smelters) had the concentrations of Cd higher than the safe level. The Sn-mining and sifting activities in Ha Thuong caused a heavy soil pollution with As, in particular the grass soil area in hamlet 7.

To protect the humans and animals in the investigated areas in the two Northern provinces of Vietnam, Thai Nguyen and Hung Yen, the soil polluted with heavy metals must be urgently remediated. The use of the land surrounding the pollution sources needs to be under the management of the local authorities in order to prevent the food chain from the penetration of heavy metals.

5. Acknowledgement

The author would like to thank the Institute of Ecology and Biological Resources at the Vietnam Academy of Science and Technology for the financial support of this study.

6. References


